Breeding bird communities along Insular Mediterranean gradients

Bruno MASSA(*) and Mario LO VALVO(°)

(*)Istituto di Entomologia Agraria, Università di Palermo, Palermo. Italy. (°)Istituto di Zoologia, Università di Palermo, Palermo. Italy.

ABSTRACT - Frequency, niche expansion, and adaptation of bird species were examined in six habitats sited along vegetational gradients in Corsica, Sicily, Crete and Cyprus (Mediterranean basin), in order to highlight possible patterns of species allocation. A peak in species diversity and total abundance is noticed in the shrubby habitats and an impoverishment in the mature ones. The number of species is proportional to the number of terrestrial breeding birds. Habitat breadth shows the highest values in the shrubby habitats of Corsica, Sicily and Cyprus, but since this pattern is slightly different at Crete, it does not seem to be correlated to isolation degree and island size. The percentage of sedentary species increases in mature habitats; an inverse correlation between habitat breadth and sedentariness seems to point out that sedentary species isolated in mature habitats should expand their ecological niche less than migrant breeders do. Succession rate follows a similar trend over the four gradients, but, according to the insularity degree, species turnover is more complete in Sicily (the largest island), less so in the other insular gradients. Correspondence analysis shows that communities are distributed along the tridimensional space of \mathbf{F}_1 - \mathbf{F}_2 - \mathbf{F}_3 planes in such a way that F_2 separates the first three habitats from the following ones. In Sicily and Corsica a higher number of species is shared in mature habitats than in Crete and Cyprus; the different sedentary/migrant species ratio in mature habitats and the more xerothermic nature of the sclerophyll forest of the eastern islands might account for the differences in bird allocation recorded on F₁-F₂-F₃ planes.

KEY WORDS: Successional gradients; Mediterranean islands; Breeding bird communities; Frequency of occurrence; Correspondence analysis

INTRODUCTION

Studies of avifauna along successional gradients show an increased bird diversity during the first vegetational stages and reduced or stable diversity in the mature ones. The increase is not monotonous along the gradient, generally peaking in the pre-climax stages (e.g. Ferry et al., 1976; Faaborg, 1980; Smith and Mac Mahon, 1981; Glowacinski and Weiner, 1983; Blondel et al., 1988; Wiens, 1989). Therefore over different

continental areas of Europe and America, bird density peaks in the mature forest (Helle and Monkkonen, 1990), in disagreement with theory predicting that density is generally correlated to habitat productivity, which in turn peaks in the pre-climax stages (Odum, 1969). In the Mediterranean area bird species diversity peaks in shrubby habitats and decrease in the mature ones (Ferry et al., 1976; Blondel et al., 1988; Lo Valvo and Massa, 1989; Massa, 1990). Contrary to the above pattern, Blondel et al. (1988) in the

Corsican shrubby habitats found both the highest species diversity and bird density. The impact of man on the Mediterranean vegetation started 10-12,000 years ago (Le Houérou, 1980), encouraging for several times the formation of garigues and different stages of the Mediterranean maquis, to the detriment of forests. Consequently forests are now relict and fragmented and hold only a restricted number of bird species, generally widespread in Europe, whereas birds inhabiting shrubby habitats are more numerous and well adapted. (Blondel et al., 1988; Blondel, 1990). On the contrary, Central European forests were permanent formations for thousands of years, this accounting for the highest diversity observed there. The aim of our research was to detect the trend of bird diversity along insular Mediterranean gradients, in order to verify the existence of the above described pattern of space partitioning, habitat use and species turnover rate over lands affected by human presence since thousands of years ago.

STUDY AREAS

Bird communities were investigated in four islands, Corsica, Sicily, Crete and Cyprus (Fig. 1), and detected along a successional gradient toward mature forest of typically Mediterranean

oaks belonging to Quercetalia ilicis association (Quercion ilicis alliance in Corsica and Sicily, Quercion calliprini alliance in Crete and Cyprus: Quezel, 1988). Each habitat is supposedly derived from the evolution of a previous one and presumably is evenly old in the four islands. The young habitats have their origin in clear felling while older habitats are of various origin, more or less in natural circumstances. Even if the vegetation structure is similar for the four islands, the compounding species are different. Some species (e.g.: Ceratonia siliqua, Olea europea, Myrtus communis, Pistacia lentiscus, etc.) are circummediterranean, others are widespread on central-western Mediterranean and absent or sporadic in the eastern area, or viceversa; Quercus ilex, for example, is episodic at Crete, absent in Cyprus and widespread in Sicily and Corsica; viceversa Q. brachyphylla (which some botanists consider to be a simple variety of Q. pubescens) is typically east-mediterranean and Q. alnifolia is endemic of Cyprus. In the island of Crete Cupressus sempervirens is common within some shrubby-arboreal vegetation, and thickets and woods in thermo- and meso-mediterranean stage are constituted by Q. brachyphylla, which is a deciduous species. Q. calliprinos of eastern Mediterranean is considered more thermophilous than *Q. ilex* (Le Houérou, 1980). The avifauna

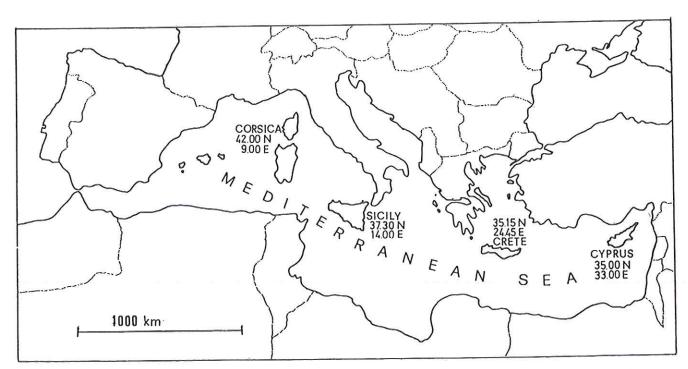


Fig. 1. Mediterranean basin and location of the four islands where bird species were censused.

 ${\it Table I}$ Botanical species and vegetation physiognomy characterizing each habitat

			ત	,Q	υ	゚゚゚	ø	Ţ	
grassland									
Corsica (C01)	-	0.5	+	0	0	0	0	0	Asphodelus sp.
Sicily (S11)	-	0.5	+	0	0	0	0	0	Asphodelus sp.
Crete (CR1)	٦	0.5	+	0	0	0	0	0	Sarcopoterium sp., Asphodelus sp.
Cyprus (CY1)	٦	0.5	+	0	0	0	0	0	Stipa iortilis, Medicago sp., Sarcopoterium sp., Asphodelus sp.
garigue									
Corsica (C02)	7	0.8	ı	1	0	0	0	0	Cistus spp.
	01	0.8	1	1	0	0	0	0	Calicotome villosa, Euphorbia sp.
	1.8	0.8	ı	1	0	0	0	0	Pistacia spp Euphorbia sp.
Cyprus (CY2)	0,	0.8	1	1	0	0	0	0	Cistus sp., Sarcopoterium sp.
low matorral (= maquis)	maqui	(S							
ದ	ო	ч	9	+	+	0	0	0	Cistus spp., Myrtus communis, Pistacia lentiscus, Calicotome sp., etc.
	2.5	7	į	+	+	0	0	0	Cistus spp., Myrtus communis, Pistacia lentiscus, Calicotome villosa, etc.
	2.5	П,	ij	+	+	0	0	0	Rhamnus oleoides. Euphorbia sp., Pistacia sp., Cupressus sempervirens, etc.
Cyprus (CY3)	2.5	1	•	+	+	0	0	0	Pistacia sp., Cistus spp., Genista sphacelata, Calicotome sp., Scutellaria cyprum, etc.
high matorral									
Corsica (C04)	4	3.5	ı	1	+	+	0	0	Arbutus unedo, Pistacia lentiscus, Erica spp., etc.
Sicily (SI4)	4	4	1	i	+	+	0	0	Arbutus unedo, Pistacia lentiscus, Myrtus communis, Calicotome villosa, Quercus ilex
Crete (CR4)	3.5	3.2	ı	1	+	+	0	0	Calicotome sp., Genista sp., Pistacia sp., Quercus calliprinos, Q. brachuphulla
Cyprus (CY4)	4	3.5	į	1	+	+	0	0	Pistacia sp., Olea europaea, Arbutus andrachne, Pinus brutia, Guercus calliprinos
thicket									
g	Ŋ	œ	ı	0	ī	1	+	+	Quercus ilex, Pistacia lentiscus, Myrtus communis, Calicotome sp.
Sicily (SI5)	Ŋ	œ	ï	0	ı	ī	+	, +	Quercus ilex. Pistacia lentiscus. Myrtus communis, Calicotome villosa
Crete (CR5)	Ŋ	7	ij	0	ï	i	+	+	Quercus calliprinos. Q. brachyphylla, Cupressus sempervirens, Chamaecutisus sp., Calicotome sp.
Cyprus (CY5)	ເດ	7.5	¢	0	ĸ	î	+	+	Quercus calliprinos. Q. alnifolia, Pinus brutia, Arbutus andrachne
mature wood									
Ü	വ	15	•	t	ı	0	+	+	Quercus ilex
	ហ រ	15	1	t	ï	0	+	+	Quercus ilex, Q. pubescens
Crete (CR6)	വവ	L L	E 1		ŧ i	0 0	+ +	+ +	Quercus brachyphylla, G. calliprinos Binis brania Ongrana calliprinos
)	2	į.	ě		>	ŀ	H	Full Supplied Company (American Company)

 $\mathbf{A} = \text{mean number of layers}$; $\mathbf{B} = \text{mean heigth of vegetation}$; $\mathbf{a} = \text{grasses (0-0.5 m)}$; $\mathbf{b} = \text{shrubs (0-0.5 m)}$; $\mathbf{c} = \text{shrubs (0.5-1 m)}$; $\mathbf{d} = \text{shrubs (1-4 m)}$; $\mathbf{e} = \text{trees (4-16 m)}$; $\mathbf{r} = \text{more than 50\% of cover}$; $\mathbf{r} = \text{less than 50\% of cover}$

was censused in six habitats; table I reports botanical species and vegetation physiognomy characterizing each habitat.

MATERIALS AND METHODS

A total of 453 point-counts were carried out: in Corsica, between Ajaccio and Porto, in 1982; in Sicily, in the area of the Madonie Mts, among the villages of Gibilmanna, Collesano, Castelbuono and Cefalù, in 1982 and 1983 (cf. Lo Valvo and Massa, 1989; Massa, 1990); on the northern coast of Crete, between Chania and Sitia, in 1983 and 1984 (by BM); in Cyprus, in the area of Larnaca, Limassol, Paphos and Polis (cf. Massa and Catalisano, 1987; Massa, 1990), in 1986 (by BM in collaboration with A. Catalisano). All the point-counts were located between 50 and 1000 m a.s.l. To remove any bias due to the census area, and to heterogeneous diversity at the edge and central areas of the same habitat (cf. Stamps et al., 1987), the point-count sites were chosen within homogeneous habitats, wide enough to allow comparisons, every effort being made to perform censuses within the habitat and not on its boundaries.

We used the EFP method (Blondel, 1975; Blondel *et al.*, 1981) to obtain the frequency of species occurrence in each habitat. The census technique was chosen for the following reasons: 1) available manpower of the censusing team; 2) mean richness yielded by this method is thought to be a reliable index of the total abundance in the community; 3) the aims of our research were to compare bird communities of similar

extensive selected habitats in different areas (cf. Blondel et al., 1981). Breeding birds were detected between 15 May and 15 July for 20 minutes in each point-count; the mere presence or absence of species in each habitat allowed us to achieve a frequency value per habitat. Point-count numbers varied from habitat to habitat and from island to island, with a minimum of 13 and a maximum of 27 per habitat (Table II). These numbers were fixed from the value of a/N, being a the number of species observed in a single point-count, and N the total number of point-counts performed within the same habitat (Ferry, 1976). In all cases a/N was lower than 0.1, thus assuring that differences in point-count numbers were not a bias source in the results (cf. Blondel et al., 1988).

By considering bird species as a variable of the six successional habitats in each island, we performed four factorial analyses of correspondences (AFC) (Benzecri, 1973; cf. Reciprocal Averaging of Hill, 1973) on the species/habitat matrix containing in its lines species frequency for each habitat. This method, which is considered appropriate for ecological analyses of gradients (Legendre and Legendre, 1976; Prodon and Lebreton, 1981; Gauch, 1986; Digby and Kempton, 1987; Blondel et al., 1988), enabled us to highlight on the F1-F2-F3 planes the point-species along the pointhabitat gradient. Values on the factorial axes are a good measurement of the species turnover along the gradient: the higher the values, the more complete is the turnover (Prodon and Lebreton, 1981). Intraspecific variance on the F1 axe corresponds to dispersal ability of each species in the different habitats of the gradient (Chessel et al., 1982); thus we used it for a

Table II

Island area (km²); total number of terrestrial regular breeding birds (A), total number of bird species censused along the whole succesional gradient (B), total (S) and mean (\$\overline{s}\$) species richness detected in the six habitats selected in the four mediterranean islands; number of point-counts in each habitat (N)

	km²	A	В				НАИ	BITATS		
					1	2	3	4	5	6
				S =	15	16	22	33	24	20
CORSICA	8721	82	43	s =	6	7	8	12	10.5	9
				N =	16	18	19	24	22	19
				S =	18	21	25	34	31	24
SICILY	25709	89	52	š =	6	6.5	7.8	11.5	11	9.5
2121-1				N =	18	19	20	27	26	20
				S =	12	18	22	27	13	11
CRETE	8222	67	36	š =	4.5	5	7	7.5	7	6
	9 1			N =	13	15	20	22	18	15
				S =	10	12	13	20	15	17
CYPRUS	9250	63	34	s =	5	5.5	5.5	9	7.5	8
~		12,070		N =	13	14	17	24	18	16

habitat-breadth estimate of the species. Starting from F₁ values of conditional means and standard deviations for each species, we built canonical graphics of habitat-breadth (Chessel et al., 1982). Moreover, in order to compare the four gradients we performed an AFC analysis on a single species/habitat matrix obtained by merging the previous ones, consisting of species frequencies for the overall 24 habitats. Community structure was stated by correlating the log of the frequencies to the species rank. The kind of regression obtained, especially the slope, consents to set and compare the distribution model for species in each habitat. Values of slopes are good indices of evenness (cf.James and Rathbun, 1981); high frequencies in the first ranks are typical of ecological models with high dominance (e.g.: log-series and brocken-stick). Steep slopes indicate low evenness, while soft slopes should match a lognormal distribution and a higher evenness (cf. May, 1975). The succession rate was measured by plotting the regression of log of turnover (TR) against time, being TR = $H'\beta/t_1$ t_0 (Glowacinski and Jarvinen 1975). H' β = 100 [H'₍₁₊₂₎-0.5(H'(1)+H'(2))], being H'(1+2) the Shannon-Wiener diversity calculated from the combined data of frequencies of two sequential habitats and H'(1) and H'(2) diversity of the two habitats separately. t_1 - t_0 is the time interval in years, estimated between a successional habitat and the following (cf. Lo Valvo and Massa, 1989).

RESULTS

Species breeding along the four gradients

Appendix 1 lists all the species detected during the censuses. Four species are present in all the habitats of the Corsican gradient (Turdus merula, T. troglodytes, Serinus citrinella and Emberiza cirlus), only two are found in Sicily (T. troglodytes and Carduelis cannabina), as in the Crete gradient (Sylvia melanocephala and Carduelis cannabina); a single species (Oenanthe cypriaca) is present along the whole Cyprus gradient. Seven species are exclusive of the Corsica gradient, 10 species of the Sicily, three and five respectively of the Crete and Cyprus gradients. Corsica and Sicily gradients share eight species, which are absent from Crete and Cyprus; conversely four species present along the Crete and Cyprus gradients are absent from Corsican and Sicilian habitats. Important differences are observable in the habitat partitioning by some species; for example Sylvia melanocephala shows an increasing habitat breadth respectively in Sicily, Corsica and Crete, in the latter also colonizing the mature habitats; its ecological substitute at Cyprus,

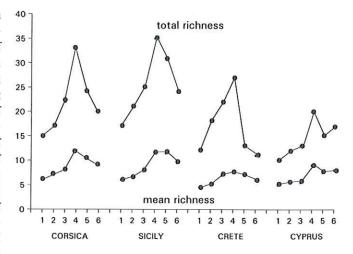


Fig. 2. Trend of total number (= total richness) and mean number (= mean richness) of bird species in the six habitats of the vegetational gradient in Corsica, Sicily, Crete and Cyprus.

S.melanothorax, is limited to the shrubby habitats, as well as S. rueppelli at Crete, where it shares the same space with S. melanocephala. Carduelis cannabina is a very euricious species along the whole gradients of Sicily and Crete, but not in Cyprus and Corsica. Another finch, Serinus citrinella, is regularly present in the Corsican gradient, from CO1 to CO6, as already pointed out by Blondel et al. (1988). Muscicapa striata shows the maximum habitat breadth in Corsica, whereas it is sporadic in Sicily (SI6) and Crete (CR4). Aegithalos caudatus is a species typical of thickets and mature woods in Sicily, but is also widespread in the high matorral of Corsica.

Species richness and evenness

All the gradients share a similar trend of mean and total species richness (Table II, Fig. 2), with a peak in the shrubby habitats and a decrease in the mature ones (habitats 5 and 6). Since mean richness as obtained by the EFP method is thought to be a reliable index of the total abundance in the community (Blondel et al., 1981), we could state that along insular gradients, shrubby habitats share both the highest bird species richness and density. On the whole, 76 species were detected, 52 of which along the Sicilian gradient, 43 in Corsica, 36 in Crete and 34 in Cyprus. These represent 61.8% of the overall terrestrial breeding birds in the four islands (n = 123) and c.55% (range: 53.7%-58.4%) of the birds breeding in each island (cf. Table II). Species number in each habitat is proportional to

Table III
Regression of species frequency model

	CORSICA	SICILY	CRETE	CYPRUS
y =	1.119-0.044x	1.139-0.050x	2.082-0.121x	2.266-0.136x
	r = -0.88	r = -0.98	r = -0.96	r = -0.96
y =	1.026-0.032x	1.102-0.049x	1.913-0.060x	1.980-0.064x
	r = -0.95	r = -0.97	r = -0.98	r = -0.89
y =	0.957-0.031x	0.894-0.026x	1.902-0.045x	1.939-0.067x
	r = -0.93	r = -0.97	r = -0.97	r = -0.94
y =	0.989-0.039x	0.899-0.030x	1.961-0.052x	1.902-0.047x
	r = -0.96	r = -0.98	r = -0.97	r = -0.96
y =	0.999-0.036x	0.924-0.031x	2.288-0.107x	2.123-0.043x
	r = -0.98	r = -0.97	r = -0.96	r = -0.91
y =	1.166-0.056x	1.019-0.038x	2.232-0111x	1.983-0.056x
	r = -0.97	r = -0.97	r = -0.96	r = -0.97
	y = y = y = y =	y = 1.119-0.044x $r = -0.88$ $y = 1.026-0.032x$ $r = -0.95$ $y = 0.957-0.031x$ $r = -0.93$ $y = 0.989-0.039x$ $r = -0.96$ $y = 0.999-0.036x$ $r = -0.98$ $y = 1.166-0.056x$	$\begin{array}{llllllllllllllllllllllllllllllllllll$	$\begin{array}{llllllllllllllllllllllllllllllllllll$

the total number detected along the whole gradient and to global richness of each island, thus higher in Sicily and increasingly lower in Corsica, Crete and Cyprus. Values for mature habitats in Sicily and Corsica result notably higher than in Crete and Cyprus.

Regression between the log of species frequency and species rank (Table III) highlights that shrubby habitats generally yield a higher evenness than pioneer and mature habitats. Small differences are noticed in the values of slopes; namely, while the highest values for Corsica and Crete are found in habitat 2 and 3 (respectively CO2, CO3 and CR2, CR3, CR4), those for Sicily and Cyprus are more shifted toward mature habitats (respectively SI3, SI4, SI5 and CY4, CY5).

Habitat-breadth

The trend of mean habitat breadth is fairly parallel along the gradients of Corsica, Sicily and Cyprus (Fig. 3). CR1 is much higher than CO1, SI1 and CY1, and consequently the trend for Crete is different from those of the other islands. Correlation of habitat breadth along each gradient versus the other ones is good and significant, with the exception of pairs involving Crete (Tab. IV). All the gradients show a similar pattern in the canonical graphics of habitat breadth (Fig. 4), with the highest values in the intermediate habitats. Only that of Crete is truncated in mature habitats.

Species sedentariness

The percentage of sedentary species shows three different patterns (Fig. 5): 1) it lies more or less at high values over the different habitats (Corsica); 2) after a slow decrease in habitats 3 and 4, it increases in 5 and 6 (Cyprus); 3) it regularly increases from 1 to 6, ranging from lowest to highest values (Sicily and especially Crete). The latter is particularly different from the Cyprus one, yielding a noteworthy number of sedentary species in mature habitats, the counterpart habitats of Cyprus showing instead the lowest

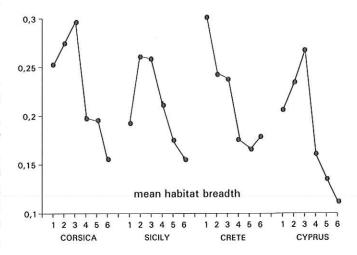


Fig. 3. Trend of mean values of habitat-breadth of breeding birds in the six habitats of the four insular gradients examined.

 ${\it Table IV}$ Coefficient of linear correlation of the parameter values of each isle versus the other ones

	Total richness	Mean richness	Habitat breadth	Sedentariness	F1 values
CO/SI	0.96 **	0.97***	0.89 **	0.29 NS	0.98 ***
CO/CR	0.71 NS	O.87**	0.67 NS	0.21 NS	0.94 **
CO/CY	0.87**	0.93 **	0.99****	0.49 NS	0.96 **
SI/CR	0.56 NS	0.84*	0.40 NS	0.98 ***	0.96 **
SI/CY	0.85*	0.91 **	0.90**	0.03 NS	0.97***
CR/CY	0.44 NS	0.68 NS	0.70 NS	0.09 NS	0.94 **
**** P < 0.0	0001 *** P < 0.001	** P < 0.02 * P < 0.0	95 NS = not sign	nificant	

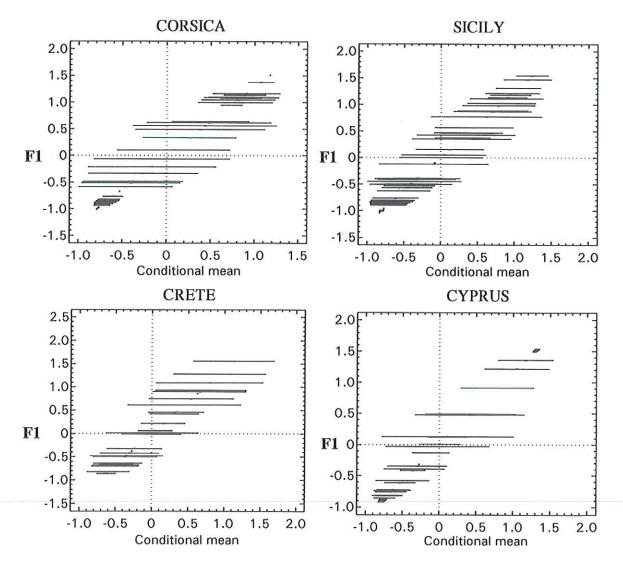


Fig. 4. Canonical graphics of habitat-breadth of breeding birds along the four mediterranean insular gradients; they were obtained by plotting conditional mean against F_1 values of species. Breadth of lines corresponds to values of conditional standard deviation. (cf. Chessel *et al.*, 1982).

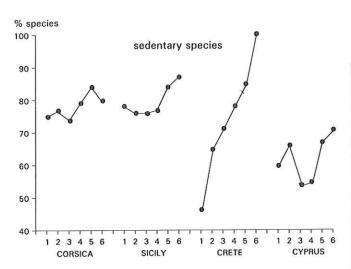


Fig. 5. Trend of percentage of sedentary species in the six habitats of the four mediterranean islands.

values. Correlation of sedentariness values for each island against those for the other islands is not significant, except for the Sicily-Crete pair (Table IV). Habitat breadth and sedentariness are inversely correlated, yet only values relative to the Crete gradient are statistically significant (SI: r = -0.61; P = NS; CO: r = -0.39; P = NS; CR: r = -0.90; P = 0.01; CY: r = -0.06; P = NS).

Succession rates

Table V reports the values of time interval, $H'\beta$ and TR, as calculated for the four gradients. TR decreases regularly in Sicily and Corsica, while it shows a little increase between CR4 and CR5 and two peaks, one between CY1 and CY2 and another between CY4 and CY5, the highest values being in CO1, SI1, CR1 and CY2. Nevertheless all the data from the four gradients merge into a single pattern, when log TR values are plotted against time: the rate of succession decreases monotonously (Fig. 6), the slope of regression being steeper and steeper from Cyprus through Corsica, Crete and Sicily.

Correspondence analysis

The projection of point-species along the point-habitats on the F_1 - F_2 plane shows a typically parabolic configuration (Fig. 7). F1 has the dominant value of variance; being related to the vegetational gradient, it exhibits an evident discriminant power between the set of the first three habitats (which always have positive values on

Tab. V
Diversity (H'β) and turnover (TR) between the following habitats of the four gradients

			НА	BIT	ATS	
		1	2	3 4	1	5 6
Time interval						
(years)		5	10	18	20	50
CORSICA	Η' β =	12.8	20.9	44.2	17.3	20.3
	TR =	2.57	2.09	2.4	0.87	0.41
SICILY	Η' β =	25.9	20	30.1	28.1	9.9
	TR =	5.18	2	1.67	1.41	0.2
CRETE	Η' β =	39.9	17.7	17.6	25.1	25.1
	TR =	8	1.77	0.98	1.25	0.5
CYPRUS	Η' β =	8.9	45.1	30.4	39.2	23.5
						0.47

 F_1) and the last three (which instead have negative values). Values of the variance as explained by the factorial axes F_1 - F_2 - F_3 are very similar in the four gradients (Corsica: F_1 =57.3%; F_2 =16.9%; F_3 =14%; total=88.2%; Sicily: F_1 =56.6%; F_2 =21.9%; F_3 =11.3%; total=89.8%; Crete: F_1 =50.4%; F_2 =21.8%; F_3 =13.8%; total=86.1%; Cyprus: F_1 =50.6%; F_2 =26.4%; F_3 =12.1%; total=89.1%). Crete shows a more regular ordination of point-species, exhibiting similar distances among the point-habitats, and a pattern of a typical continuum. Point-species of Corsica and Sicily are more scattered and dis-

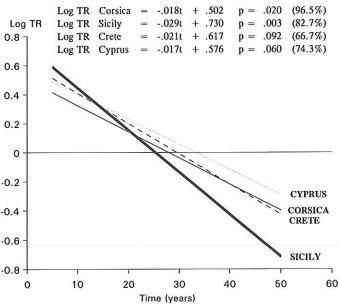


Fig. 6. Succession rate in the four gradients, calculated plotting the values of logTR against time in years. TR = $H'\beta/t_1$ - t_0 (cf. Table V).

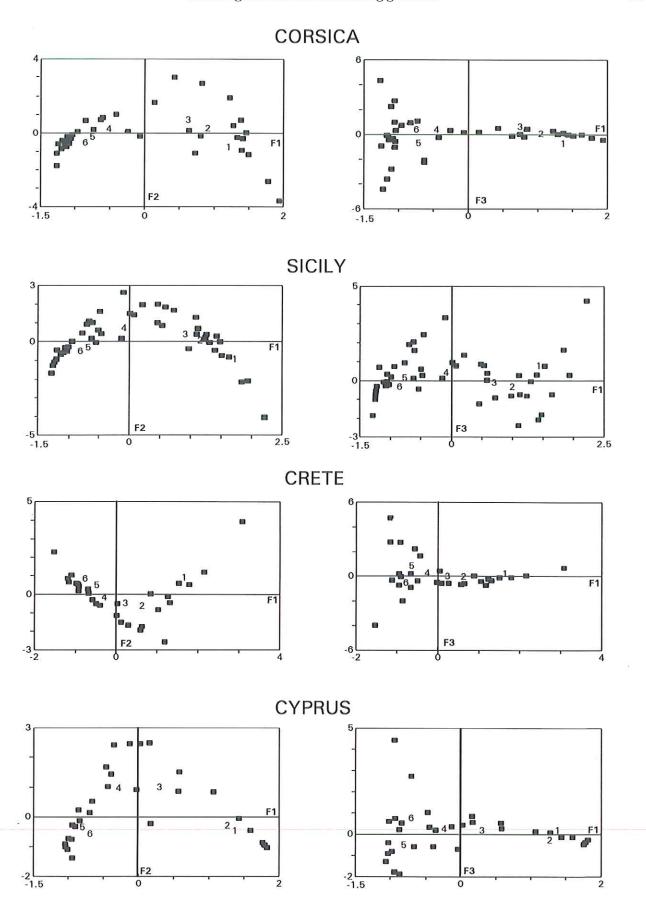


Fig. 7. Projection of the bird species (squares) and habitats (1 to 6) on the F_1 - F_2 (left) and F_1 - F_3 (right) planes of the correspondence analysis in the four insular gradients.

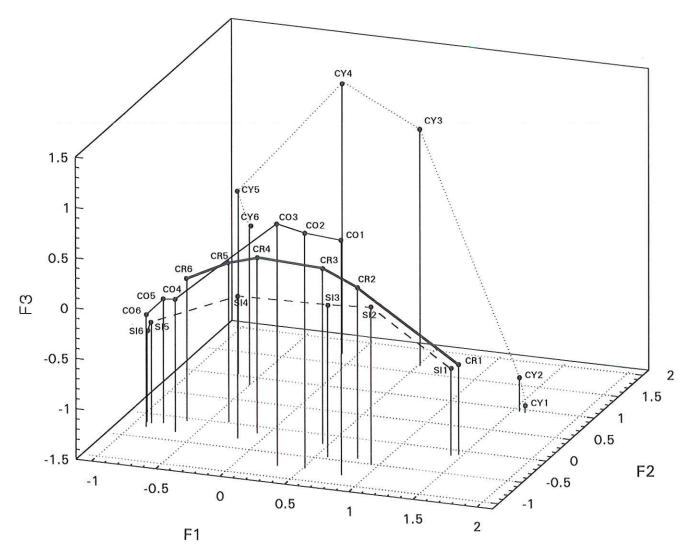


Fig. 8. Ordination on the F_1 - F_2 - F_3 planes of the six habitats, obtained by correspondence analysis of a cumulative matrix of the four insular gradients.

tributed, in such a way that it is evident a sharp passage between the habitats 1-2-3 and 4-5-6, probably due to a real dissimilarity between the two sets. Finally, the point-species of the Cyprus gradient are much confused and overlapped; in this case three sets of point-habitats are identifiable, namely 1-2, 3-4 and 5-6, dependent on a greater separation by relative bird communities. Distribution of point-species on F_1 - F_3 plane is more or less sygmoidal-like in all the gradients. The projection of 24 point-habitats of the F₁-F₂-F₃ planes (Fig. 8), drawn from a cumulative matrix of the four gradients, still shows the typical parabolic pattern of the gradients, which converge on the mature habitats, as already observed by Blondel et al. (1988: Fig. 5). The F_1 - F_2 - F_3 axes account only for 47% of the total variance $(F_1=20.6\%; F_2=15.1\%; F_3=11.3\%)$, being F_1 inertia scarcely dominant. As in the previous analysis, F_1 plane separates definitely 1-2-3 from 4-5-6 habitat-sets. Nevertheless, what seems more evident in Fig. 8 is the distance of the Cyprus gradient from the other three; in particular the point-species of this gradient seem well distinct on the F_1 - F_3 plane, resulting in a separate pattern, with respect to the others.

DISCUSSION

Table IV summarizes correlations of the set of parameters of each island versus the other ones. Many differences between pairs of trends may be observed, in particular in those involving Crete; in spite of their geographical proximity, the Crete-Cyprus pair shows the least similar trends. Altogether, our results lead us to con-

clude that, even if by multivariate analysis we recognize a pattern of bird allocation rather similar for the four insular gradients, important divergences exist in some parameters of their avian communities. Incidentally, this suggests to be cautious in comparing islands/continent bird data, particularly for the eastern Mediterranean islands.

We try to compare breeding avian communities in similar habitats, as censused in a definite time in different geographical areas. Our censuses satisfy three of the four recommendations by Helle and Monkkonen (1985) for this type of analysis: 1) habitats were selected in such a way that the time interval between analogous of different gradients is of similar length; 2) avifauna was censused in different sites of the same habitat in order to quantify the intra-habitat variability; 3) the census method used was effective (cf. Blondel et al., 1981; Martin, 1983; Bart and Klosiewski, 1989; Petty and Avery, 1990). However, our censuses fail to satisfy the fourth point, since our study takes into account only the breeding season, the influence of temporal variation of communities being unknown. Owing to the lack of winter climatic stress, Mediterranean ecosystems can supply vital resources to large numbers of wintering migrants (Telleria et al., 1988), the majority of which take advantage, in shrubs, of the peaking fruit abundance, which occurs two months later than in Central Europe (Herrera, 1982; Jordano, 1985; Snow and Snow, 1988). The flow of wintering individuals in local populations could modify the community structure of an area and the resource exploitation within a limited time period. Temporal effects may shift communities along the equilibriumnon equilibrium continuum: thus long-term dynamics of communities in a single site might yield very different conclusions from short-term studies of similar communities in different sites scattered over a large area (Jarvinen and Haila, 1981; Wiens, 1984, 1989; Wiens et al., 1986).

Nevertheless, our data on some of the islands were gathered during different years and it is possible that by looking at relative distribution patterns, the effect of annual differences may be minimized. Along two of the four gradients birds were censused during two successive years, respectively 1982-83 (Sicily) and 1983-84 (Crete). Results were divided into two sets and analysed separately, but we failed to find any

significant difference between pairs of values (range of correlation between frequency values: 0.91-0.98 in the six habitats); besides, censuses along the Sicilian gradient were repeated in 1989-90, with no important differences being found with respect to those of 1982-83 (range of correlations: 0.92-0.96). Finally, our results for the Corsican gradient are satisfactorily in agreement with those provided by Blondel *et al.* (1988) for the same island; this encourages us to suppose that no significant short-time variation occurs within the bird communities along the gradients here investigated.

Some authors (Ferry et al., 1976; Martin, 1982; Lo Valvo and Massa, 1989) have pointed out the immigration of species from mature habitats of Corsica and Sicily toward shrubby habitats, not balanced by the opposite movement, this accounting for the higher habitat breadth of shrubby species. Our results confirm a general shifting of species toward shrubby and pioneer habitats in insular gradients, so much that all the habitat breadth trends show the highest values in the first three or four habitats, especially in the low shrubby ones. Crete holds just 11 species in mature habitats, 100% of which are sedentary; they are all (with the exception of Certhia brachydactyla) able to colonize shrubby habitats, thus increasing habitat breadth values. As a matter of fact, 10 out of 11 species (91%) of CR6 also inhabit CR4 and CR5. This accounts for the significant inverse correlation found at Crete between sedentariness and habitat breadth, as well as the truncated pattern of canonical graphic of habitat breadth (Fig. 4). The trend of graphics for other gradients instead should depend on a lower percentage of species from mature habitats being able to colonize shrubby ones (namely 11/20 = 55% in Corsica, 9/24 = 37.5% in Sicily, and 5/17 = 29% at Cyprus).

Species of the four gradients do not show a habitate expansion related to their geographical isolation: Sicily, separated from the Italian continent by a sound of just 3 Kms, has indeed higher habitat breadth values than Crete and Cyprus, respectively 100 and 70 Kms away from Greece and Turkey. Corsica (80 Kms away from the closest continent) and Sicily show a moderate impoverishment in the mature habitats, while Crete and Cyprus show a dramatic species diminution in the same habitats. Division of ecological space in the four gradients should occur in a

different way, and the sole geographical factors do not account for their insularity attributes. Bird turnover rate in the four gradients seems a good index of insularity: more sloped regressions should mean a more complete species turnover rate from habitat 1 to 6. Insular conditions generally have the attribute of an incomplete turnover, some species colonizing the whole gradient, namely habitats 1 to 6, and more species than in continental gradients being present from 1 to 5 or 2 to 6 (Blondel, 1986; Blondel et al., 1988). It resulted more complete in Sicily (the least isolated island and of largest size), less at Crete, Corsica and Cyprus. The slope value recorded in the Sicilian turnover (0.29) lies within the minimum continental ones reported by Glowacinski and Jarvinen (1975).

Sicily revealed more similarities with bird communities of Corsican gradient than with Crete and Cyprus ones. Sicily and Corsica are much different in their paleogeographical, environmental and isolation aspects, and their convergence could be explained by botanical affinity. This is greater between the sets of the habitats of Corsica and Sicily than with those investigated at Crete and Cyprus (note that Sicily and Corsica vegetation alliance, Q. ilicis, is different from and less thermophilous than that of Crete and Cyprus, Q. calliprini: Le Houérou, 1980); there are also some differences in the botanical aspects of vegetational gradient of Crete and Cyprus, Q. brachyphylla being absent from mature habitats sampled at Cyprus. Consequently some differences in the community assemblage could depend just on the botanical differences in the gradients (cf. Rotenberry, 1985 vs. Cody, 1981). Greater and older human impact on mature habitats of eastern Mediterranean islands, with respect to western ones could account for the higher bird poorness in mature habitats of Crete and Cyprus, with respect to those of Corsica and Sicily. Moreover eastern Mediterranean sclerophyll forest zone has a more xerothermic nature, merging in its drier borders with semiarid and less productive Irano-Turanian steppe formations (Naveh and Lieberman, 1984). Poorness of mature habitats of Crete and Cyprus might depend on their location at the most eastern part of the Mediterranean area, so much influenced by the climate of the Asian steppe areas, which in turn possibly affected the distinctiveness of bird allocation by their communities.

Acknowledgements

We wish to thank K. De Smet, T. Mingozzi, J.L. Telleria and J.A. Wiens for their useful comments and suggestions on an earlier draft of the manuscript. We also express our appreciation to A. Catalisano for her assistance to B.M. in the field work at Cyprus. This paper was financially supported by Ministero Università e Ricerca Scientifica e Tecnologica (60% funds, 1992) and by Consiglio Nazionale delle Ricerche. We dedicate this paper to Jacques Blondel, for his seminal work on the bird communities along the gradients.

REFERENCES

- BART, J., and KLOSIEWSKI, S.P., 1989. Use of presence-absence to measure changes in avian density. *J. Wildl. Manage.*, **53**: 847-852.
- Benzecri, J.P., 1973. L'analyse des données. II. Dunod. Paris. 619 pp.
- BLONDEL, J., 1975. L'analyse des peuplements d'oiseaux, éléments d'un diagnostic écologique. I. La mèthode des echantillonnages frequentiels progressifs (E.F.P.). Terre et Vie, 29: 533-589.
- BLONDEL, J., 1986. Biogéographie évolutive. Masson. Paris.
- BLONDEL, J., 1990. Biogeography and history of forest bird faunas in the Mediterranean zone. *In: Biogeography and Ecology of forest bird communities.* Keast, A.(Ed.). SPB Academic Publ.. The Hague, Nederlands. Pp.95-107
- BLONDEL, J., FERRY, C., and FROCHOT, B., 1981. Point counts with unlimited distance. *In: Estimating numbers of terrestrial birds*. Ralph, C.J. and Scott, J.M. (Eds.). *Studies in Avian Biology* **6**: 414-420
- BLONDEL, J., CHESSEL, D., and FROCHOT, B., 1988. Bird species impoverishment, niche expansion, and density inflation in Mediterranean island habitats. *Ecology*, 69: 1899-1917.
- CHESSEL, D., LEBRETON, J.D., and PRODON, R., 1982. Mésures symmétriques d'amplitude d'habitat et de diversité intrachantillon dans un tableau espèces-relevés: cas d'un gradient simple. C.R. Acad. Sc. Paris, 295: 83-88.
- CODY, M.L., 1981. Habitat selection in birds: the roles of vegetation structure, competitors and productivity. BioScience, 31: 107-113.
- DIGBY, P.G.N., and KEMPTON, R.A., 1987. Multivariate analysis of ecological communities. Chapman and Hall, London, 206 pp.
- FAABORG, J., 1980. Potential uses and abuses of diversity concepts in wildlife management. *Trans. Missouri Acad. Sciences*, 41-49.
- FERRY, C., 1976. Un test facile pour savoir si la richesse mesurée d'un peuplement se rapproche de sa richesse réelle. *Le Jean le Blanc*, **15**: 21-28.
- FERRY, C., BLONDEL, J., and FROCHOT, B., 1976. Plant successional stage and avifaunal structure on an island. *Proc.* 16th Int. Orn. Congr., 643-653.
- GAUCH, H.G. jr., 1986. Multivariate analysis in Community ecology. Cambridge University Press, 298 pp.
- GLOWACINSKI, Z., and JARVINEN, O., 1975. Rate of secondary succession in forest bird communities. *Ornis Scand.*, 6: 33-40.
- GLOWACINSKI, Z. and WEINER, J., 1983. Successional trends in the energetics of forest bird communities. Holarctic Ecology, 6: 305-314.
- Helle, P., and Monkkonen, M., 1985. Measuring turnover rates in secondary succession in European forest bird communities. *Ornis Scand.*, 16: 173-180.
- Helle, P., and Monkkonen, M., 1990. Forest successions and bird communities: theoretical aspects and practical implications. *In: Biogeography and Ecology of forest bird communities*. Keast, A. (Ed.). SPB Academic Publ.. The Hague, Nederlands, pp. 299-318
- HERRERA, C.M., 1982. Seasonal variation in the quality of fruits and diffuse coevolution between plants and avian dispersers. *Ecology*, 63: 773-785.

- HILL, M.O., 1973. Reciprocal averaging: an eigenvector method of ordination. J. Ecol., 61: 237-249.
- JAMES, F.C., and RATHBUN, S., 1981. Rarefaction, relative abundance, and diversity of avian communities. Auk, 98: 785-800.
- JARVINEN, O., and HAILA, Y., 1984. Assembly of land bird communities on northern islands: a quantitative analysis of insular impoverishment. In: Ecological communities: conceptual issues and the evidence. Strong, D.R. jr., Simberloff, D., Abele, L.G. and Thistle, A.B. (Eds.). Princeton Univ.Press, Princeton, N.J., pp. 138-147.
- JORDANO, P., 1985. El ciclo annual de los pesseriformes frugivoros en el matorral mediterraneo del sur de Espana: importancia de su invernada y variaciones interanuales. Ardeola, 32: 69-94.
- LEGENDRE, L. and LEGENDRE, P., 1979. Ecologie numerique. 2 vols., Masson, Paris, France.
- Lo Valvo, M., and Massa, B., 1989. Les communautés d'oiseaux nicheurs dans des successions à Chêne vert Quercus ilex en Sicile et en Corse. Alauda, 57: 308-318.
- MARTIN, J.L., 1982. Le diagnostic de la compensation de densité dans les peuplements insulaires d'oiseaux par la méthode des Echantillonnages Frequentiels Progressifs (E.F.P.). Acta Oecologica, 4: 167-179.
- MARTIN, J.L., 1983. L'infiltration des oiseaux forestiers dans les milieux buissonnants de Corse. Rev. Ecol., 36: 397-419.
- MASSA, B., 1990. Bird communities along a secondary succession in Mediterranean and Canary Islands. Atti Conv. Lincei, 85: 215-231.
- MASSA, B., and CATALISANO, A., 1987. Considerations on the species richness detected along an ecological succession of Cyprus. Annual Report of Cyprus. Orn. Soc.,: 61-64.
- May, R.M., 1975. Patterns of species abundance and diversity. *In: Ecology and Evolution of communities*. Cody, M.L. and Diamond, J.M. (Eds.). Belknap Press of Harvard Univ. Press. Cambridge, Mass., USA, pp. 81-120
- NAVEH, Z., and LIEBERMAN, A.S., 1984. Landscape Ecology. Springer Verlag, New York, New York, USA.
- Petty, S.J., and Avery, M.I., 1990. Forest Bird Communities. Forestry Commission Occasional Paper 26, Edinburgh, UK.
- Prodon, R., and Lebreton, J.D., 1981. Breeding avifauna of a Mediterranean succession: the holm oak and cork oak series in the eastern Pyrenees, 1. Analysis and modelling of the structure gradient. *Oikos*, **37**: 21-38.
- Quezel, P., 1988. Esquisse phytogéographique de la végétation climacique potentielle des grandes iles méditerranéennes. Bull. Ecologie, 19: 121-127.
- ROTENBERRY, J.T., 1985. The role of habitat in avian community composition: physiognomy or floristics? *Oecologia*, **67**: 213-217.
- SMITH, K.G., and Mac Mahon, J.A., 1981. Bird communities along a montane sere: community structure and energetics. *Auk*, **98**: 8-28.
- Snow, B., and Snow, D., 1988. Birds and berries. T. & A.D. Poyser, Calton, England.
- Stamps, J.A., Buechner, M., and Krishnan, V.V., 1987. The effects of habitat geometry on territorial defense costs: intruder pressure in bounded habitats. *Amer. Zool.*, 27: 307-325.

Telleria, J.L., Suarez, F., and Santos ,I., 1988. Bird communities of the Iberian shrubsteppes. *Holarct. Ecol.*, **11**: 171-177.

Wiens, J.A., 1984. On understanding a non-equilibrium world: myth and reality in community patterns and processes. *In: Ecological communities: conceptual issues and the evidence*. Strong, D.R.jr., Simberloff, D., Abele, L.G. and Thistle, A.B. (Eds.). Princeton Univ. Press, Princeton, N.J., pp. 439-457. WIENS, J.A., 1989. The Ecology of bird communities. 1. Foundations and patterns. 2. Processes and variations. Cambridge Univ. Press, Cambridge, UK.

Wiens, J.A., Addicott, J.F., Case, T.J., and Diamond, J., 1986.
Overview: the importance of spatial and temporal scale in ecological investigations. *In: Community Ecology*. Diamond J. and Case T.J. (Eds.), Harper & Row, New York, New York, USA: 145-153.

Received Jannuary 1993, Accepted September 1993

Address and correspondance:

Bruno Massa Ist. Entomologia Agraria Viale delle Scienze, 13 I-90128 PALERMO ITALY

Alectoris gracea Alectoris criukar collinus Alectoris criukar collin			Τ	CC	R	SIC	CA.	ah		S	SIC	ZIL	Y	ah	CI	RE'	ГE		ah	(CY	PR	US		ah
Alectoris Fulda		s						T	1	2				0.04	-				1	1				1	
Fremcoltunes francoltunis Structure		s	1	2	3	4		0.14	1					0.01											
Premieration francollinis Streephopelia turiur Columic continues Streephopelia turiur Contracts genrulus								300000 0						1	2	3	4		0.09	2	2 3	4	5		0.50
Columbe polanthus								8 88	1							- 1000	1000			2	2 3		-		
Sirepopela turtur			1	2			_		1				_ /-		1										0.00
Catcuits centoris														100000000000000000000000000000000000000											
Coracias garmulus		18/5/5															5	ř.	0.00			4	5	6	0.04
Pipupa epigus		1000000				4	об	0.01				4	5 6	0.06	1										
Priodeles mignor		15000				4	6	0.02				4	5.6	0.08											
Jynk forquilla		73373373													1							4		ь	0.04
Calantered brachytadestyla Calantered bra	Jynx torquilla	1 6										- 85	0 0	0.00						1					
Calaridate brachytactified Calaridate brachytactified Calaridate orbores Section Calaridate orbores Section Sect	Melanocorypha calandra	s							1					0.00						1 2					0.00
Lullula arborea Alauda arborea Ala	Calandrella brachydactyla																		0.00						
Alauda arvensis			١.	_	_							20								1 2					
Anthus campestris S 1 2 3 4 5 6 0.06 12 3 4 5 6 0.06 12 3 4 5 6 0.06 12 3 4 5 6 0.08 12 3 4 5 6 0.02 0.00		100000	1			4			1	2	3	4		0.24	2	3	4		0.10				- 1	6	0.00
Troglodytes troglodytes S 1 2 3 4 5 6 0.00 1 2 3 4 5 6 0.06 1 2 3 4 5 6 0.06 1 2 3 4 5 6 0.08 1 2 3 4 5 6 0.00 1 3 4 5 6 0.00 1 3 4		2000	1						١,		2			0.10	1.0	0	4		0.04						
Ertifiacius rubecula		1000				4 !	5.6					4 5	5.6					G					= 1		0.00
Luscinta megarhynchos Theoreticurus ochrunos Saxicola torquata Saxicola torqua		1000	1	-					1	4					4	J	4 5	О	0.25				D (2	0.00
Phoenteurus colurors Saxcola torquata Connenthe operation Connenthe in this panica Connenthe connenthe in this panica Connenthe connenthe in this panica Connenthe in this panica Connenthe in this panica Connenthe connenthe in this panica Connenthe in this panica Connenthe connenthe in this panica Connenthe in this panica Connenthe connenthe in this panica Connenthe in this panica Connenthe connenthe in this panica Connenthe in this panica Connenthe connenthe Connenthe connenthe in this panica Connenthe connenthe Connenthe connenthe in this panica Connenthe connenthe		100		2						2											3	4			0.06
Saxiola torquata Concunting community Saxiola torquate Concunting community Concunting community Concunting community Concurrence Concurrenc	Phoenicurus ochruros								1 :	2	3	4	25		1						J	T			0.00
Cenanthe oenanthe m					3	4						4		0.28					0.55					1	
Demanthe cupiriaca	Oenanthe below		1					0.00	1 :	2	3			0.08			4		0.15						
Monticola solitarius Turdus menula Turdus visciorus S 1 2 3 4 5 6 0.53 2 3 4 5 6 0.00		100000000000000000000000000000000000000	4	0	0	4 .		0.00							12	3			0.31						
Turdus menula		A100 A100 A100 A100 A100 A100 A100 A100	1	2	3 '	4 :	0 6	0.80	١,,,	,	0			0.00					0.00					1	
Turdus viscisorus Cettita cettit S S 3 4 0.29 3 4 0.00 1 3 4 5 6 0.00 1 3 4 5 6 0.00 1 3 4 5 6 0.00 1 3 4 5 6 0.00 1 3 4 5 6 0.05 5 5 5 0.00 1 3 4 5 5 0.05 5 5 0.00 1 3 4 5 5 0.05 5 0.00 1 3 4 5 5 0.05 5 0.00 1 3 4 5 5 0.05 5 0.00 1 3 4 5 5 0.05 5 0.00 1 3 4 5 5 0.05		258	1	2	3	4 -	5.6	0.53	1 2)	3	4 5	5.6		2	2	1 =	C							
Cetta cett		223	ै	2	υ.	т .	, 0	0.00	1 1	4	J .					0	4 5	О	0.20						
Cisticola funcidis	Cettia cetti				3 4	4		0.29		1	3 4		, 0		2	3	4	- 1	0.09						
Hippotais pallida			1		88				1 2								•	- 1		12					0.00
Sylpia sarda		m															4 5	-				4	5 6		
Siglia conspicillata			1			2								567 88753											
Siylota cantillans Siylota		10274			3 4	4		0.29	١.,	377					100.00	181		- 1		1					i
Siglicia melanocephala Siglicia melanocephala Siglicia melanochorox Siglicia melanochorox Siglicia melanochorox Siglicia articapilla Siglicia arti		3355			9	4 5		0.20							12	3			0.14						
Sigluta melanathroax		100000													1.0	9	1 =		0.41						
Siglicia rueppelli		1 . Sec. 1		2	0 -	1 0	•	0.42	1 4		, .	1 0	,	0.22	12	0 '	4 3	0	0.41		Q	1	5		0.16
Syluica communits Syluica continents Syluica communits Syluica community Syl																3 4	4		0.06		J	4	J	1	0.16
Phylloscopus collubita		m								:	3 4	1		0.15		~	•		0.00						
Regulus ignicapilius		s			4	1 5	6	0.01		:	3 4	1 5	6			4	4		0.00						
Musicicapa striata		1																							
Aegithalos caudatus		1. (298)										5					a	- 1	NE SEE						
Parus ater Parus caeruleus S		8 1,725		1								_				4	1		0.00			4		1	0.00
Parus caeruleus		3000										b	'	0.00										١.	
Parus major	The state of the s										Δ	15	6	0.06	1 1	2 /	15	6	0.10				5 6	' '	0.00
Sitta europaea	Parus major									9												4	5 6	، ا	0.04
Certhia brachydactyla		s													3	_		٦,	0.00				0 0	1	0.04
Lanius collurio	Certhia familiaris					5	6	0.00						M 0 8080											i
Lanius senator		233.0						0.00				5	6	0.00				6	0.00				6	(0.00
Lanius nubicus		100000		;	5 4	ł		0.29			1,2	ú		0.00		0			0.5						
Carrulus glandarius		100000000000000000000000000000000000000							2		4	ŀ		0.29	12	3 4	ł		0.40		0	4	- ~	1.	0.17
Pica pica s 5 0.00 4 5 6 0.12 4 5 6 0.05 4 5 6 0.05 4 5 6 0.00 4 5 6 0.12 4 5 6 0.05 4 5 6 0.06 0.06 0.06 0.06 0.00 4 5 6 0.01 0.00 4 5 6 0.01 0.00 4 5 6 0.01 0.00 4 5 6 0.01 0.00 4 5 6 0.00					4	. 5	6	0.02			4	5	6	0.07							3				
Corvus corone							_	0.02											- 1				JO	1,	0.00
Passer italiae						5		0.00								4	1 5	6	0.05			4	5	1	0.06
Passer domesticus		777143			4	ŀ							-2	SAMSATTI									.	1	
Passer hispaniolensis Passer montanus Passer montanus S Passer montanus S Passer montanus S Passer montanus S Petronia petronia Fringilla coelebs S Serinus serinus S Serinus serinus S Serinus citrinella S S Serinus citrinella S S S S S S S S S S S S S S S S S S		5500									35.90	E DOV		2 50144								4	6	1	0.03
Petronia petronia																								1	
Fringilla coelebs s 4 5 6 0.02 4 5 6 0.06 3 4 5 6 0.11 4 5 6 0.02 Serinus serinus s 4 5 0.01 3 4 5 6 0.34 4 5 0.03 4 5 6 0.02 Serinus citrinella s 1 2 3 4 5 6 0.66 0.66 0.09 3 4 6 0.11 3 4 5 6 0.02 Carduelis chloris s 3 4 5 6 0.33 4 5 6 0.09 3 4 6 0.11 3 4 5 6 0.13 Carduelis carduelis s 1 2 3 4 5 0.50 1 2 3 4 5 6 0.45 2 3 4 5 6 0.61 0.17 1 2 3 4 6 0.55 Carduelis cannabina s 1 2 3 4 5 6 0.55 2 3 4 5 6 0.56 12 3 4 5 6 0.61 1 2 3 4 0.37 Emberiza cia s 1 2 3 4 5 6 0.55 2 3 4 5 6 0.46 2 4 0.17 Emberiza melanocephala m m 2 3 4 0.11 0.00 3 4 0.07		263							1.0			5													
Serinus serinus Serinu		2020			1		6	0.02	12	C		E	6		9	9	F .	6	011			4		١,	000
Serinus citrinella		200								9					3										
Carduelis chloris			1	2 3						·	-1	J	٦	0.04		12	· U		0.03			+ :	JO	1,	0.02
Carduelis carduelis S 1 2 3 4 5 0.50 1 2 3 4 5 0.45 0.45 0.45 0.45 0.17 1 2 3 4 6 0.55	Carduelis chloris			3	3 4	5					4	5	6	0.09		3 4	į i	6	0.11		3	4 !	5 6	1	0.13
Carduells cannabina				2 3	3 4	5		0.50			4	5			2 3	3 4	5								
Emberiza cirlus		10000	1	2 3	3 4	1000	_							0.56	123	3 4	5								
Emberiza melanocephala m 3 4 0.08 Emberiza hortulana m 2 3 4 0.11 2 0.00 3 4 0.07 Militaria caesia m 10.00 1		0.07%	1	2 3	3 4	5	6	0.55					6		2									1	
Emberiza hortulana m 2 3 4 0.11 Emberiza caesia m m 2 0.00 3 4 0.07									2	3	4	8		0.13							_				
Emberiza caesia m 2 0.00 3 4 0.07		(A.S.)													0.4) A			0.13		3	4		10	0.08
Militaria seleculus 1 0 0 4 0 10 1 0 0 4 0 0 1		45.000					- 1									4	E .				2	Л		1	0.07
0.10		200000	1 :	2 3	4			0.18	12	3	4			0.23								T			
								34487F	1,7(3,612)					(5) (5) (5) (5) (6)	155			18	03		_			1	

APPENDIX - List of species censused along the four gradients. Numbers reported in each line refer to habitat in which the species was detected. \mathbf{s} = sedentary species; \mathbf{m} = migrant breeding species; \mathbf{ah} = mean value of habitat breadth.

	1
	*
	- <u>x</u>
	*
	· ·
	Ŷ